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31 Abstract

32 Food, energy, and water (FEW) systems have been recognized as an issue of critical global 33 importance. Understanding the mechanisms that govern the FEW nexus is essential to develop 34 solutions and avoid humanitarian crises of displacement, famine, and disease. The US and China 35 are the world's leading economies. Although the two nations are shaped by fundamentally 36 different political and economic systems, they share FEW trajectories in several complementary 37 ways. These realities place the US and China in unique positions to engage in problem definition, 38 dialog, actions, and transdisciplinary convergence of research to achieve productive solutions 39 addressing FEW challenges. By comparing the nexus and functions of the FEW systems in the two 40 nations, this perspective aims to facilitate collaborative innovations that reduce disciplinary siloes, 41 mitigate FEW challenges, and enhance environmental sustainability. The review of experiences 42 and challenges facing the US and China also offers valuable insights for other nations seeking to 43 achieve sustainable development goals.

44

45 Introduction

Humanity is poised on the precipice of global crises entailing growing needs for food, energy, and water (FEW) resources in the face of increasing climate change stress and dynamic demographic and socioeconomic transitions. A United Nations Environment Programme report¹ on achieving sustainable development goals highlights the lack of progress and even "negative change" for several indicators related to the availability and security of adequate FEW resources. Breakdowns in FEW systems can lead to human displacement, disease, and suffering. Understanding the dynamics of the evolution of the FEW nexus and the mechanisms that govern

future FEW relationships is essential to mitigate potential humanitarian disasters associated with
 the lack of clean water, accessible energy, and nutritious food.²

55 China and the United States (US) are the world's leading economies, with the highest levels 56 of production and consumption of FEW resources and associated effects on the environment and 57 society. These realities place China and the US in unique positions to engage in problem definition, 58 transdisciplinary convergence of research, and complementary actions to achieve productive 59 solutions addressing FEW challenges on a global scale. Yet, geopolitical and ideological barriers are being raised that increasingly restrict US-China research collaborations.^{3, 4} The aim of this 60 61 perspective is to highlight differences and similarities between the two nations in FEW nexus 62 challenges and opportunities and to identify steps that could facilitate collaborative solutions.

63 Developing solutions for FEW security and transitioning nexus knowledge to sustainable 64 actions are facilitated by nurturing transdisciplinary collaborations that advance understanding of the multiscale impacts of FEW resource consumption.⁵ FEW systems are characterized by 65 66 complex inter-system interactions, socioeconomic impacts, and challenges of grand stressors (e.g., climate change and population growth).⁶ The interactions (e.g., resource supply and demand) 67 68 underlying FEW systems cut across all sectors of society and occur at scales ranging from garden 69 plots and individual households, to regional and national production and consumption, and to 70 international trade and supply chains. The interconnectedness of production, transport, and 71 consumption of various goods means that any solution towards the sustainable exploitation of 72 FEW resources must consider their joint needs, opportunities, constraints, and feedbacks at local 73 to global scales.

74

76 Existing Dissimilarity

77 The US and China differ in their flows of FEW production, transport, and consumption (Figs. 78 1 and 2). The FEW flows have led to significant differences in the nexus and impacts of FEW 79 systems in the two nations. The most obvious difference is that the interdependences of FEW 80 consumption are relatively imbalanced relative to those in the US. For instance, coal is the primary 81 energy source for industry in China, whereas natural gas, petroleum, and renewable energy are 82 widely used along with coal for industry development. With regard to the energy-food nexus, 83 modern agriculture is deeply dependent on energy, and energy determines the level of agricultural mechanization, farming intensification, and the use of fertilizers, pesticides, and plastic films.⁷⁻⁹ 84 85 Agriculture-related energy use in the US is not illustrated in Fig. 1; however, the US Department 86 of Agriculture (USDA) reported that the agricultural sector demanded 1.74% of total US primary energy consumption in 2014.¹⁰ In China, a lower amount of $\sim 1\%$ of the total national energy 87 88 consumption is directed towards the agricultural sector (Fig. 2). Key differences between the two 89 nations are the extent of mechanization and the rate of change in employing advanced agricultural 90 technologies. While mechanized systems have dominated US agriculture for decades, China is 91 currently in a widespread and rapid transition phase, and the contributions of mechanized farming to agricultural output increased from about 44% in 2007 to over 70% in 2019.^{11, 12} The production 92 93 of most staple crop (wheat, rice, and maize) now relies on mechanized systems, and the numbers 94 of middle to large tractors has shown a 20% annual increase.¹² This mechanization has substantially accelerated energy consumption in the agricultural sector in China.^{8, 13} For example, 95 96 the total power of agricultural machinery approximately doubled from 523 million kW in 2000 to 1,026 million kW in 2012.¹⁴ 97

98 Focusing on the water-food nexus, agriculture is the second largest water consumer in the US 99 and the largest user in China. Agricultural irrigation accounts for 37% of US total freshwater 100 consumption, with about half the volume coming from surface water and half from groundwater 101 resources (Fig. 1). Irrigation has by far the largest demand on US groundwater (68% of all 102 groundwater withdrawals), with public water supply systems a distant second (18% of 103 withdrawals). Agricultural water use in the US is primarily dedicated to maize (Zea mays L.) and 104 soybean (Glvcine max) in the western plains, cotton (Gossypium) and rice (Oryza sativa) in the 105 Mississippi valley, forage crops in western states, and specialty crops in California's Central Valley.¹⁵ 106

107 In comparison, food production systems are the primary sink for water in China, accounting 108 for 63% of all water utilization (Fig. 2). Increasing meat consumption and intensification of animal 109 production systems will exacerbate the demand for freshwater.¹⁶ Agriculture impacts water quality 110 through both soil erosion and contamination as a consequence of increasing reliance on pesticides and chemical fertilizers.^{17, 18} Therefore, importing water-intensive food commodities (e.g., corn 111 112 and soybean) from the US not only is favorable for protecting the quality of limited water resources 113 but also can save a significant volume of water for industrial and domestic needs in China. The 114 role of "virtual water" via agricultural trade obviously provides opportunities for improving the 115 supply of affordable food while conserving resources in water-scarce regions.^{19, 20}

The energy→water nexus is derived from energy consumption during the extraction and allocation of water, wastewater treatment, and water heating for domestic and industrial uses.^{7, 21-}
Moving water, including groundwater pumping and surface water transfers over long distances, is energy-intensive. Moreover, to ensure access to clean and potable water, additional energy-intensive processes, such as desalinization, filtration, and wastewater treatment, have been

implemented.²¹ The water-energy nexus is intensified by increasing water use in the energy 121 sector.²⁴ For instance, water withdrawals for cooling thermal electric power plants represent the 122 123 largest single demand sector in the US, or 41% of all water use (Fig. 1). The total water use in the energy sector was estimated at 2.22×10^{11} m³ of water withdrawals in 2015, mainly for fossil fuel 124 125 extraction and processing. While 56% of surface freshwater withdrawals in the US served cooling 126 purposes in power plants, demand from this sector will decline as older coal and nuclear power 127 plants are being retired and replaced with natural gas and renewable energy sources. In addition, increased energy efficiency will reduce the demand on energy and water resources.^{25, 26} In contrast, 128 the water→energy nexus in China appears less favorable to conservation in the near term. The coal 129 industry in China consumed 7.4×10^8 m³ of water in 2015, which is 40% more than water required 130 if natural gas would be the energy carrier.²⁷ Reduction of coal-based energy could thus lessen 131 water scarcity in China.²⁸ While Fig. 2 does not show water withdrawals for energy directly, one 132 133 recent analysis found that water withdrawals for energy production in China have decreased from 4.34×10^{10} m³ in 2002 to 1.56×10^{10} m³ in 2015²⁹. This decline is partly attributed to China's active 134 135 participation in global energy trade, which allows more petroleum and natural gas to meet the 136 country's energy needs and replace traditional coal to some extent. Of course, international 137 cooperation and agreements are a prerequisite for international energy trade that can help to alleviate the water scarcity of import country (e.g., China).²⁷ However, energy trade might worsen 138 the water scarcity in the energy export countries (e.g., Kazakhstan and Saudi Arabia).³⁰ A mutually 139 140 favorable trade should thus be importing energy from water-abundant country. If a water-141 sustainable energy trade is impossible between two countries, the trade should involve a third 142 country to balance the loss of "virtual" water. For instance, the energy export country that lacks 143 water should trades in water-consuming foods (e.g., cereals) from another country that has no

water scarcity problem. Overall, energy use for water currently dominates the energy-water nexus in China since China's food security heavily relies on exploration of water resources (Fig. 2). In comparison, energy use for water and water use for energy are well balanced in the US. (Fig. 1). This difference implies that a balance between energy and water consumptions is critical for sustainable development and that crop production in accordance with regional water availability might contribute to the mitigation of local energy-water conflict.³¹⁻³³

150

151 Asynchronous Similarities

152 Despite historical, cultural, and political differences, the US and China share common 153 objectives and challenges regarding FEW systems. Balancing demands for safe and nutritious food, 154 secure and sufficient energy, and clean water among people and sectors, while also confronting 155 the uncertainties surrounding climate change, is a shared challenge. Furthermore, while the timing 156 is distinct, the two nations have similar patterns of production and consumption of FEW resources. 157 While ramping up food production, both China and the US are attempting to increase water-use 158 efficiency and to improve water quality by adopting advanced irrigation systems in fields and 159 developing rainwater-harvesting and wastewater reuse technologies. Sustenance of high living 160 standards and thriving rural communities while embracing more sustainable yet cost-competitive agricultural production systems are a common core goal.^{34, 35} Cereals and oilseeds are the top two 161 162 food classes produced in both nations (Figs. 1-3). The total food production in the US and in China was similar $(8.45 \times 10^8 \text{ versus } 9.70 \times 10^8 \text{ tonnes in } 2015$, respectively) (Figs. 1, 2). While China 163 164 consumed 99% of the food domestically, the US exported 19%.

165 China has a thousand-year history of feeding the world's largest population through labor-166 intensive agriculture.³⁴ China's rapid movement toward a developed economy has overturned the peasant farming model and threatens to hollow out China's rural areas following a pattern similar to what was observed in rural America, where big businesses and confined animal feeding operations (CAFOs) replaced small family farms.³⁶ In addition, increasing affluence and rapid urbanization have substantially affected food consumption patterns in China.³⁷ The growing urban population is shifting toward higher levels of consumption of meats and dairy products.^{37, 38} This new domestic market has resulted in a rapid, often uneven, expansion of animal production in China since the 1990ies (Fig. 3).

174 In the US, animal and food production systems have matured over decades, with persistent 175 improvements in yields and efficiency. Agricultural technologies (e.g., GPS-assisted vehicle 176 guidance systems, yield and soil mapping, and advanced equipment) and agricultural policies 177 through farm subsidy program have driven increases in both quantity and quality, leading to consolidation of farmland and the growth of corporate farming.^{39,40} Meanwhile, increased food 178 consumption in the US is mainly driven by the population growth.³⁴ The structure of food 179 180 production in recent decades has been relatively stable with a 1% average annual increase in 181 quantity since 1999, in part due to relatively stable but higher levels of income and food 182 consumption. The US enjoys an abundance of food products provisioned by domestic and 183 imported sources.

In both the US and China, food abundance is unequally distributed within society.⁴¹ For instance, about 11% of US households were food-insecure in 2019, meaning access to enough food for an active, healthy life for all household members was not assured.⁴² Relatively wealthy and diverse consumers drive increasing agricultural imports to the US (e.g., a 6% increase in 2018).⁴³ Agricultural imports in China (which rose by 3% in 2018) are increasing to secure basic food needs of its 1.4 billion people and for animal feed to meet growing demands of urban consumers.⁴¹

190 In terms of energy, both nations rely asynchronously on fossil fuels, with coal and oil supplying 191 \sim 70% of the energy needs (Figs. 1-3). Fig. 1 shows that total energy consumption in the US was 2.25×10^9 tonnes (8.92×10^{16} Btu) in 2015 and relied heavily on petroleum and natural gas. The US 192 193 energy supply was used primarily for electric power production, transportation, and industry. 194 Approximately 51%, 26%, and 14% of renewable energy were consumed by electric power plant, 195 industry, and transportation, respectively. Petroleum was mainly used for transportation, and 196 natural gas was consumed primarily by electric power plants and industry. Similarly, Fig. 2 shows the total energy consumption in China was 4.30×10^9 tonnes in 2015 with heavy reliance on coal. 197 198 This energy was consumed primarily by industry. Specifically, all of the crude oil and most of the 199 coal and renewable electric energy were used for industry (Fig. 2). Overall, Fig. 3 shows that China 200 surpassed the US in energy consumption in 2010. The sources of consumed energy in China lacked 201 diversity prior to 2000. Coal was the dominant fossil fuel (contributing more than 80% energy 202 needs) followed by oil fuel. The source diversity of energy consumption increased after 2000 and 203 accelerated after 2010. For instance, the consumption of natural gas, nuclear power, and renewable 204 energy (e.g., hydropower) increased, though coal and oil fuels still occupied large proportions.

205 The differences in consumption and structure of energy between the two nations are closely linked to their levels of urbanization and industrialization.^{37, 44, 45} The urban population of China 206 207 increased from 172 million in 1978 to 771 million in 2015, with an average annual increase of 16 208 million. As a result, the urbanization rate (i.e., percentage of urban population in total population) increased from 18% to 56%, with an average annual absolute increase by 1%.⁴⁶ The pace of 209 210 China's urbanization has significantly accelerated since 1996, with the average annual increase 211 maintained at a rate of over 1.3% during the period from 1996 to 2015. This urbanization rate was 212 4.5 times higher than that during 1949-1978 and 1.9 times higher than that during 1978-1996.⁴⁶

213 This demographic change has substantially increased the energy consumption in the urban areas. 214 In addition, China has been experiencing an industrial transition from primary industry towards 215 secondary and tertiary industries. This transition since 2000 has led to a rapid increase in energy consumption and facilitated the diversification of energy composition.⁴⁵ In contrast, during the 216 217 same period, the energy consumption in the US remained relatively stable. Urbanization in the US 218 increased from 79.6% in 2000 to 81.7% in 2015, with an average annual absolute increase of 0.14%.⁴⁷ This very slow increase in urbanization, along with an upgrade of economic structure 219 220 towards higher value activities in production (e.g., advanced engineering and pharmaceutical 221 development), flattened the energy consumption.⁴⁵ Currently, similar trajectories of energy 222 consumption and urbanization as occurred in China and the US are occurring in other developing (e.g., Indonesia) and developed countries (e.g., Germany), respectively.⁴⁸ Predictions based on the 223 224 Food-Demand Model (FDM), total energy use is expected to increase by 78% in developing countries and by ~1% in developed countries during the period from 2015 to 2050.⁴⁹ 225

226 Total water withdrawals are very similar between the two nations and range from 400 to 620 227 billion m³/year (Fig. 3). Water consumption illustrates that water use in both nations relies on 228 surface water by $\sim 80\%$ and groundwater by $\sim 20\%$. Fig. 1 shows that the total water consumption in the US was 4.44×10¹¹ m³ in 2015, with most coming from fresh surface water. Fresh surface 229 230 water accounted for almost all of the water used in thermoelectric energy production and most of 231 the water used in industrial production and aquaculture. Fig. 2 shows that the total water consumption in China was 6.10×10¹¹ m³ in 2015, with most being fresh surface water. Most of the 232 233 water was used for agriculture (Fig. 2). Use of water for agriculture in China has slowly declined since 1992.⁵⁰ Specifically, the increase in agricultural water use slowed down from an acceleration 234 rate of $+8.33 \text{ km}^3/\text{y}^2$ in 1965-1975 to $+3.09 \text{ km}^3/\text{y}^2$ in 1975-1992 to $-1.99 \text{ km}^3/\text{y}^2$ afterward.⁵⁰ The 235

236 responses of farmers to the changes in local climate, market condition, and irrigation subsidies are 237 key drivers of increases in irrigation efficiency. Urbanization also played a role in increasing water 238 consumption by domestic sectors as a result of water-intensive lifestyle that urban people have.⁵¹ 239 In the US, agricultural production mainly relies on large-scale farming systems, which are mostly 240 coupled with advanced water-conserving technologies for high efficient irrigation. As a result, the 241 US used less water for agriculture than China. Water resources have thus been saved to support 242 industrial needs, such as raw materials extraction and conversion processes and energy generation 243 (e.g., thermoelectric, hydropower, and nuclear power). Afterwards, the wastewater is reclaimed and returned to river systems.^{30, 52} 244

245 In past decades, both nations have emphasized investments in massive-scale infrastructure 246 projects for securing water and energy resources. For instance, China South-North Water Transfer 247 Project (SNWTP) consists of three routes to move vast volumes of water from regions in the south to support municipalities and agriculture in the dry north and central regions.^{53, 54} The project's 248 249 East Route and Middle Route have been in operation since 2013 and 2014, respectively, while the West Route is still in planning.⁵⁵ The East Route transports nearly 9 km³ of water per year from 250 251 the Yangtze River Basin to the Yellow River Basin over distances of up to 2,000 km. The water 252 transfer via this route could rise to as much as 14.8 km³ per year by 2030. The East Route pumps 253 water uphill over 65 m through a large-scale system of pumping stations from the Yangtze River 254 to its destinations. The system of pumping stations consumed electricity at a rate of 0.15 kWh/m³, ⁵⁶ which means a total of 2.35 billion kWh of energy was consumed to transfer 15.5 billion m³ 255 256 water during the period from November 2013 to May 2017. This energy production embodied consumption of 7.4 million m³ of virtual water during the operation period, which indicates 257 258 transferring 100 m³ of water consumes 0.05 m³ of water due to the electricity consumption. Hence

the East Route of SNWTP will consume 1.35 billion kWh of energy and 4.6 million m³ of virtual 259 water in order to transfer 7.3 billion m³ water by 2030. The Middle Route can carry up to 9.5 260 261 km³/year of water over 1,246 km from the Danjiangkou Reservoir on the Han River (a tributary to 262 the Yangtze) and cross Henan and Hebei provinces before reaching its destination in Beijing and 263 Tianjin for residential, agriculture, and industrial uses. A future plan for the Middle Route is to 264 eventually transfer 14 km³ of water per year. The West Route crosses the Qinghai-Tibetan Plateau, 265 transferring 17 km³ of water per year from the headwaters of the Yangtze River Basin to the 266 headwater of the Yellow River Basin.⁵⁴ Similar public works projects in the US were implemented 267 to move water from the Rocky Mountains to California and arid southwestern states.⁵⁷ The 268 inception of California State Water Project (CSWP) was in 1960, and carries up to 5 km³ of water per year over 1,126 km from northern to southern regions in California.⁵⁸ This mega-project 269 270 supplies water to more than 27 million people ranging from the San Francisco Bay Area through 271 the San Joaquin Valley, to the Central Coast and southern California. The transferred water 272 irrigates 3,035 km² of farmland, mainly in the San Joaquin Valley and generates an average of 273 6,500 GWh of hydroelectricity annually.

274

275 Disconnected Efforts

Despite sharing similar objectives and challenges related to the FEW nexus, major barriers exist in interdisciplinary research as well as international research efforts between China and the US. Historically siloed subsystems, which focus solely on specific missions while ignoring system connectivity, stymie development of approaches involving international, transdisciplinary teams. Recent policies and trade disputes have further complicated efforts to develop scientific research collaborations.^{3, 4} Neither the US nor Chinese societies have yet produced a skilled workforce

282 capable of effectively addressing complex sustainability issues. Awareness of the FEW nexus is 283 generally low, which limits the political will to invest in research. Useful engineering solutions in the peer-reviewed literature may be overlooked because of lackluster support from stakeholders.⁵⁹ 284 285 This problem is mainly caused by a lack of collaboration between researcher and stakeholder. As 286 a result, some scientific investigations are not open to stakeholders, and some research findings 287 are not actionable at stakeholder level. Failure to incorporate human behaviors and cultural 288 differences among various interest groups creates disparity between the FEW research community 289 and stakeholders, including farmers and consumers, corporations, nongovernmental organizations, and policymakers.⁶⁰ 290

291 An additional challenge is the need to collect consistent and comparable data and effectively 292 coordinate and synthesize datasets generated in different FEW disciplines. Although creation and 293 analyses of large data sets are making great progress in both countries, those advances have not often been applied to FEW systems.^{61, 62} Also, while the complexity of FEW-system interactions 294 295 requires extensive collaborations and big data science approaches, the perceived or actual inequity of credit sharing can discourage researchers from engaging in the process.⁶³ The reward and 296 297 assessment systems that credit scholarly accomplishments (e.g., publications) in both countries are 298 discipline-specific and differ for researchers in engineering, business, natural science, and social science disciplines.^{64, 65} Disparities also exist among tenure-track faculty, research faculty, 299 extension faculty, government scientists, and private sector researchers and consultants.^{65, 66} At 300 301 many academic institutions, performance measures rely on outdated metrics and rarely embrace collaborative efforts outside the home discipline.⁶⁶ The traditional thinking is counter-productive 302 303 for enticing researchers to engage in FEW nexus investigations.

305 Steps toward Solutions

306 Successful implementation of innovative solutions requires inclusion of scientists, engineers, 307 and other stakeholders in the entire process of co-producing knowledge to enhance the 308 sustainability of FEW systems. These stakeholders need to be engaged from the start, with 309 identification of the problem, through the process to identify and test solutions, and to implement selected innovations and technologies.⁶⁷ Thus, a first step toward constructive collaborations on 310 311 FEW nexus issues is to link people from traditionally disconnected disciplines and places through 312 a common vision and specific goals. Teaching curricula, organizational structures, and rewards for 313 advancement need to be revised to give researchers deserved credit for contributions to large, 314 transdisciplinary team efforts and to ensure that outcomes are understood, agreed upon, and 315 communicated. Better understanding of the benefits derived from such research partnerships could 316 help catalyze US participation.⁵

317 The second step is to develop universal metrics that can effectively monitor progress in terms 318 of human welfare and sustainable outcomes. Well-defined measures of success can help guide both implementation and future FEW research agendas.⁵⁹ The United Nations' Sustainable 319 320 Development Goals (SDGs) include indicators and targets designed to address global goals such as eradicating poverty and hunger by 2030.^{1, 68} The interconnectedness of these goals means that 321 322 all countries must cooperate in implementing and monitoring global goals, sharing knowledge, 323 and sharing technology. SDGs to eliminate poverty and hunger and provide clean water and 324 sufficient energy to all require transformation in food production and delivery systems that focus 325 on efficiency, nutritional quality, and sustainable production systems.

The third step toward more sustainable FEW systems is to provide decision makers with clear, balanced choices based on technical, economic, and life cycle analyses of integrated systems.

328 Since international relations and disturbances affect supply and demand of FEW resources, system 329 models should incorporate sociopolitical influences, global trade considerations, demographics, 330 and changes in disturbance regimes. For example, growers around the globe are learning the value 331 of timely market information and careful planning so that harvests can retain a competitive 332 advantage within global markets and respond to supply disruptions caused by extreme weather or 333 other calamities around the world. Therefore, development of smart data-based prediction system 334 for regional and global trade of FEW resources is critical to buffer temporary shortfalls, respond to natural disasters, and enhance resilience.⁶⁹ A shared vision developed from data analysis could 335 336 identify win-win goals while acknowledging disparate national or local needs. For instance, while 337 some nations may emphasize economic growth and/or sociopolitical stability (e.g., the U.S. and 338 China), other nations (e.g., Switzerland, Sweden, and Norway) pay more attentions on environmental sustainability while ensuring energy security and equity.⁷⁰ 339

340 The fourth step is to create effective mechanisms for integrating different approaches and to 341 validate the adaptively optimized schemes. To date, no single theoretical or empirical model exists 342 that satisfies the needs of FEW nexus researchers and stakeholders. This gap is because each FEW 343 system has different boundary conditions, stakeholders, and socioeconomic and environmental 344 settings. However, comparison and infusion of different approaches can help accelerate 345 demonstrations, learning, and development of a FEW research agenda. The agenda should evolve 346 in response to changing needs and conditions, while contributing to demonstrations of how science 347 applied to human-environmental systems can reduce FEW crises and enhance benefits in a 348 complex world. To achieve these goals, integrated efforts that engage many stakeholders should 349 incorporate cultural differences in order to be able to address local and global threats and to 350 implement changes that might avert disasters.⁶⁷

351 The fifth step is to develop shared standards and communication platforms that ensure 352 implementation of sustainable FEW solutions among different countries. For instance, artificial 353 intelligence systems could be employed to more quickly and effective share data and knowledge. 354 Reliable interfaces are required to transfer and transform information so that it can be used to meet 355 the diverse needs of different communities. A virtual platform can be established that allows 356 different stakeholders to proactively collaborate on common tasks related to specific FEW nexus 357 solutions. Technological developments in the US and China are already interconnected. Many 358 electronic devices designed in the US are assembled in China, and most Chinese technology firms 359 rely on foreign suppliers. Thus, instead of thinking of the US and China as competitors, FEW goals 360 are better achieved if the two countries work in partnership.⁷¹ Such cooperation has implications 361 for making progress toward sustainable development goals for food security, affordable renewable 362 energy, and access to clean water.

363

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552 Fig. 1 Sankey diagrams of food, energy, and water consumption flows in the US in 2015. The

553 width of flows reflects the proportion between the sources and sectors of consumption. Data

- 554 sources: OECD-FAO Agricultural Outlook 2015-2024, US Geological Survey (USGS), and US
- 555 Energy Information Administration (EIA).



- **Fig. 2** Sankey diagrams of food, energy, and water consumption flows in China in 2015. The
- 558 width of flows reflects the proportion between the sources and sectors of consumption. The
- diesel et al. source in the energy includes diesel, petroleum, fuel oil, kerosene, and natural gas.
- 560 Since the proportion data for the sources corresponding to the sectors of water consumption in
- 561 China is not available, the flow relationship between the sources and sectors of water
- 562 consumption is not shown. Data sources: OECD-FAO Agricultural Outlook 2015-2024 and
- 563 China Statistical Yearbook by National Bureau of Statistics.



565 Fig. 3 Food production, energy consumption, and water withdrawals in China (left column) and 566 the US (right column). The x-axes for water differ from those for energy and food due to the 567 limitation of data. The distribution of water withdrawals in China was based on the database in 568 1980, 1985, 1990, 1993, 2000, 2007, 2012, and 2015. The withdrawals in the US was based on 569 the database in 1990, 1992, 2000, 2002, 2005, 2007, 2010, and 2014. Data sources of food production, water withdrawals, and energy consumption of both countries are United Nations 570 571 Food and Agriculture Organization (UNFAO) in 2016 and 2017, Our World in Data (https://ourworldindata.org/water-use-stress) and AQUASTAT of UNFAO, and BP Statistical 572

573 Review of World Energy in 2016, respectively.

TOC Art (JPEG Image)



578 Biography of Professor Jie Zhuang

579 Dr. Jie Zhuang is a professor in Department of Biosystems 580 Engineering and Soil Science and Center for Environmental 581 Biotechnology at the University of Tennessee, Knoxville, Tennessee, 582 USA. As one of founders, he has served as the deputy director of the 583 China-US Joint Research Center for Ecosystem and Environmental



584 Change since 2007. He created a US-China 100-PhD Program in the areas of Environment, Energy 585 and Food in 2014 and has recruited more than 40 students for PhD study in the United States. 586 Currently, with the financial support of National Science Foundation of the United States, Dr. 587 Zhuang leads a project aiming to develop a global research network of food-energy-water nexus 588 for supporting urban sustainability. This transdisciplinary project involves researchers, students, 589 stakeholders, and policymakers of many countries of the world. Over the past three decades, Dr. 590 Zhuang has worked on many challenging research projects in the United States, Japan, and China. 591 His specific research focuses on the fate and transport of contaminants (pathogens, radionuclides, 592 colloids, organic chemicals, and munitions constituents), soil carbon management, soil hydrology, 593 and plant-water relations.